

Clarifications on Select Topics for the External Peer Review of EPA's Inorganic Arsenic (iAs) IRIS Assessment

Presentation for the Science Advisory Board Inorganic Arsenic Review Panel

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EPA held a public planning and scoping workshop



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SEPA United States Environmental Protection Agency Elements Performed by IRIS

IRIS assessments are systematic reviews of publicly available scientific studies on environmental agents, with 2 goals:

1.Qualitative \rightarrow the <u>nature</u> of hazardous effects

2.Quantitative \rightarrow the <u>concentrations</u> associated with effect induction



Actions by other EPA Programs and Regions



Public Opportunities during the iAs Assessment Development

- Nov 2012: EPA released Problem Formulation Materials for public comment
- Jan 2013: EPA hosted an iAs workshop to discuss planning and scoping
- Mar-Aug 2013: EPA hosted Arsenic Webinar Series
- June 2014: EPA hosted a public science meeting on preliminary materials (e.g., assessment plan, literature search strategy, literature search results, study evaluations)
- May 2019: EPA released an *Updated Problem Formulation and Systematic Review Protocol* for the iAs IRIS Assessment for public comment
- July 2019: NAS hosted public meeting to evaluate Updated Problem Formulation and Systematic Review Protocol for the iAs IRIS Assessment
- Oct 2023: EPA released the draft assessment for public comment
- Dec 2023: EPA released BMD Model Code and BMD Modeling Results for public comment



Major Elements in Scoping and Problem Formulation



Focus on Human Data for Dose-Response

- NRC (2013) concluded that human data are expected to be the basis for dose-response analyses
- Using human data avoids uncertainty related to species extrapolation and relevance, and it allows for observations in human variability
- Human data are preferred as recommended in guidelines: EPA guidelines, technical, and methodological documents state that human data is preferred over animal data when epidemiologic studies of sufficient quality are available (EPA Cancer Guidelines, external NRC/NASEM reports on iAs)

EPA United States Environmental Protection Agency Mode of Action (MOA) Analyses

- 2019 IRIS iAs Protocol: Considerable efforts were undertaken to conduct MOA analyses to determine whether the available MOA evidence can inform dose-response of health outcomes
 - Approach presented in current draft assessment supported during 2019 NASEM peer review
 - The majority of the committee [10 out of 11 members] agrees that <u>MOA information will</u> <u>not contribute directly to determining the shape of the dose-response curve based</u> <u>on epidemiological data</u>..."
- 2023 iAs Draft Appendix A of Updated Problem Formulation and Protocol
 - Analysis of modes of action common to multiple health effects
 - Reactive oxygen species (ROS) generation and oxidative stress responses, As(III) binding to thiol groups and inhibition of key enzymes, As(V) inhibition of oxidative phosphorylation, cell cycling and damage repair impairment, epigenetics, endocrine disruption, cytotoxicity and regenerative proliferation
 - ~5726 studies screened, 191 studies summarized in appendix A
 - Case study using bladder cancer to address feasibility of using MOA and mechanistic data to inform dose-response
 - Introduction of uncertainty due to in vitro studies conducted at high concentrations, assumptions of
 applicability of in vitro model systems to human response, ability to extrapolate in vitro concentrations to
 human exposure levels, and difficulty from extrapolating from rodent studies

MOA Approach Adheres to Cancer Guidelines

nental Protection

- The Cancer Guidelines state that MOA analyses are used to "address the question of human relevance of animal tumor responses; to address the differences in anticipated responses among humans...; and as the basis of decisions about the anticipated shape of the dose response relationship.":
 - <u>Human relevance</u>: iAs is a known carcinogen with a large amount of high-quality epidemiological evidence with carcinogenic risk to humans established by multiple organizations (NRC, ATSDR, IARC, WHO.)
 - <u>Human variability</u>: extensive information on risk modifiers is available in numerous epidemiological studies
 - <u>Dose-response</u>: there are abundant epidemiological studies of low-level exposure to inorganic arsenic

Health Effects Evaluation and Conclusions

Non-cancer effects

- Circulatory system (*evidence demonstrates*)
- Diabetes (evidence demonstrates)
- Pregnancy and birth outcomes (*evidence indicates*)
- Neurodevelopmental effects (*evidence indicates*)

Well-established cancer effects not re-evaluated*

- Bladder Cancer: human carcinogen
- Lung cancer: *human carcinogen*
- Focus was on dose-response

* Classification of *human carcinogen* for these cancer effects (NRC, ATSDR, IARC, WHO) was adopted

Other outcomes considered to be out of scope

- Skin cancer, skin lesions, and immune effects (RRB comparison)
- renal cancer, liver cancer, and nonmalignant respiratory (moderate evidence and no benefit-cost need)

iAs Dose-Response Methodology **Environmental Protection** Agency

• Dose-response approach described in Allen et al. (2020a,b) and Appendix C



Data Adjustment/Pre-analysis Adjustment for Covariates: **Effective Counts** Meta-Regression: Dose-Response Analysis **Define Hierarchical** Structure **Define Parameter** Priors Parameter **Updating:** Pooled Dose Effect **Extrapolation to Target Population** Increased Lifetime Probability of Effect (Risk) as a **Function of Dose**



- Study-specific exposure metrics were converted into a unified daily intake metric $\binom{\mu g}{kg-day}$
- Exposure factors necessary for conversions identified from multiple sources
- For example, for a study reporting cumulative exposure $({}^{\mu g}/_L \times years)$, conversion was carried out as follows:

 $dose = DI + f \times (WCR \times WE) + (1 - f) \times (WCR \times LE)$

DI = dietary intake (daily µg/kg) f = fraction of time spent consuming water WCR = water consumption rate (L/kg) WE = water concentration (µg/L) LE = low end water concentration (µg/L)





- Probability distributions were inferred for each conversion factor with parameters based on the reported means and standard deviations
- Monte Carlo (MC) analysis was conducted, sampling the assumed distributions N times, where N = study specific dose group size (truncated at N = 1000)
- Individual daily intake values averaged across samples
- This process was repeated 1000 times to generate a MC distribution where 5th percentile, mean, and 95th percentile values were used in subsequent dose-response analysis



TA Tronmental Protection Dose Conversions – Urinary Biomarker Studies

- The El-Masri and Kenyon (2008, 2018) PBPK models provides the basis for assuming a 1:1 ratio between ug iAs/kg-day oral intake and μg total As/kg-day urinary excretion
- EPA assumed that urinary levels come from iAs intake, from drinking water consumption that varies across exposure groups and a study-specific estimate of dietary intake that is constant across exposure groups

A Dose Conversions – Urinary Biomarker Studies

• EPA assumed that, for a given individual, the following equation and coefficients reported by Forni Ogna et al. (2015) and el Masri and Kenyon (2018) can be used:

$$\mu mol \frac{creatine}{kg - day} = \beta_0 + \beta_1 \times sex + \beta_2 \times BMI + \beta_3 \times age + \beta_4 \times age^2$$

- EPA estimated an oral ug iAs/kg-day dose for each individual in each study dose group using the above equation and study relevant distributions for height, body weight, age, sex and total arsenic urinary excretion
- The average of the simulated population ug iAs/kg-day exposure group estimates were estimated via MC sampling and then used as doses for the primary dose-response analyses

Bayesian Dose-Response Meta-Analysis (DRMA)

- Purpose is to combine data from multiple cohort and case-control studies
- EPA assumes the prospective likelihood is given by a logistic equation, with arsenic intake, X, as the explanatory variable

$$logit{\Pr(D = 1|X) = \alpha^* + \beta(X)}$$

- Allows estimation of prospective likelihood from case-control studies (which are retrospective by definition) and thus inclusion with cohort studies in a DRMA
- Logistic model flexible enough to represent nonlinear "sigmoidal" dose-response relationships expected at a population level for toxicants with widely differing individual sensitivities (e.g., due to human heterogeneity and/or multiple iAs MOAs) (NRC, 2006, 2009, 2014)
- This type of sigmoidal shape is possible at the population level "even if the doseresponse relationship has a clear threshold in a single rodent species or cell line" (NRC, 2014)

Bayesian Dose-Response Meta-Analysis (DRMA)

- Priors for hierarchical Bayesian modeling:
 - Normal distribution for individual studies
 - Gamma for pooled estimate (does not allow negative values, consistent with causal HI determination)
- Output of modeling is a set of studyspecific logistic slopes and a pooled estimate of the logistic slope

Table 4-1. Prior parameter values for meta-regressions

Parameter	Prior distribution
β(i)ª	Normal (β_mean, β_sigma)
β_mean	Gamma (a = 0.52, b = 1.12)
β_sigma	Half-Cauchy (scale = 5)

 ${}^{a}\beta(i)$ is the dose coefficient for data set i.

Charge Question 4b

	b	0.9097	0.005	0.3925	0.1828	0.6354	0.8945	1.1689	1.7216
	vlambda[1]	0.9019	0.0061	0.5295	0.1868	0.5089	0.8036	1.1805	2.2069
Lin et al. (2018)	vlambda[2]	1.4529	0.0098	0.844	0.3011	0.8322	1.3016	1.8961	3.5238
<u>,</u>	vlambda[3]	0.6385	0.0044	0.3728	0.1306	0.3657	0.5684	0.8387	1.5605
	OR_RR[1]	1	NaN	0	1	1	1	1	1
	OR_RR[2]	1.2532	0.0015	0.121	1.0454	1.167	1.2429	1.3286	1.5197
	OR_RR[3]	1.7357	0.0052	0.4076	1.1113	1.4433	1.6763	1.9641	2.7027
Pooled	lβ_mean	0.3138	0.0026	0.1956	0.0048	0.1654	0.3056	0.4407	0.7342
β_5	sigma	0.5804	0.0029	0.2118	0.2886	0.4355	0.5397	0.6831	1.09



Credible intervals are highlighted.





- Age stratified morbidity and mortality data used to calculate lifetime probability of disease at assumed background dose
- Individual study and pooled logistic slopes then used to estimate lifetime probability of disease over a range of hypothetical doses
- Extra lifetime risk calculated as $ER = \frac{P(d) - P(0)}{1 - P(0)}$

			b						
		5th	mean	95th		extra ris	sk; d = 0.13	3 ug/kg-d	
						5th	mean	95th	
Chen 2010	1	0.0411746	0.0752772	0.106897	1	1.02E-04	1.87E-04	2.66E-04	
Steinmaus 2013	2	0.3238836	0.5148644	0.70916	2	8.11E-04	1.30E-03	1.79E-03	
Wu 2013	3	0.7902154	1.0393566	1.296354	3	2.00E-03	2.65E-03	3.33E-03	
Bates 1995	4	-0.723699	0.3279311	1.400362	4	-1.76E-03	8.21E-04	3.61E-03	
Steinmaus 2003	5	-0.826559	-0.076506	0.604969	5	-2.01E-03	-1.89E-04	1.53E-03	
Bates 2004	6	-0.327208	-0.175284	-0.04004	6	-8.05E-04	-4.33E-04	-9.92E-05	
Meliker 2010	7	-0.551266	0.2049776	0.920782	7	-1.35E-03	5.11E-04	2.34E-03	
Baris 2016	8	-0.083815	0.6510162	1.454691	8	-2.07E-04	1.64E-03	3.76E-03	
Chang 2016	9	0.0296468	0.1151488	0.19821	9	7.36E-05	2.87E-04	4.94E-04	
Huang 2018	10	0.2789166	0.5908449	0.917179	10	6.97E-04	1.49E-03	2.33E-03	
Lin 2018	11	0.2928053	0.9096531	1.588773	11	7.32E-04	2.31E-03	4.12E-03	
	12				12	0.00E+00	0.00E+00	0.00E+00	
	pooled	0.0160786	0.3137993	0.650794	pooled	3.99E-05	7.85E-04	1.64E-03	
					*10,000	0.399112	7.852797	16.43872	



Cancer Dose-Response: Bladder cancer Environmental Protection

• Carcinogenic to Humans

Agency

 Eleven epidemiologic studies in diverse populations used in DRMA

Table 4-3. Pooled meta-regression estimates of extra lifetime bladder cancer incidence risk (per 10,000) at various average daily iAs doses using MLE doses^{a, b}

Extra lifetime		Average daily inorganic arsenic dose (µg/kg-day) ^b											
risk estimates (per 10,000)ª	0	0.02	0.0365 ^b	0.075	0.13	0.185	0.24	0.57	1.12				
5 th percentile	0	0.06	0.11	0.23	0.40	0.57	0.74	1.76	3.47				
Mean	0	1.19	2.17	4.49	7.85	11.27	14.75	36.87	79.04				
95 th percentile	0	2.44	4.48	9.32	16.44	23.81	31.44	83.24	197.26				

^aExtra lifetime risks are presented as mean risk/10,000 with 5%–95% probabilities based on mean, 5% and 95% estimates of dose-response slopes.

^bDoses used in EPA modeling. U.S. daily background dose is estimated at 0.0365 µg/kg, 0.02 µg/kg from diet, $0.0165 \,\mu\text{g/kg}$ from water and $0 \,\mu\text{g/kg}$ from air (see Section 4.3.4).



Lifetime Cancer Slope Factor (CSF): 1.27 $\times 10^{-2} (\mu g/kg-d)^{-1}$

Charge Questions 5c,6

Cancer Dose-Response: Lung cancer nvironmental Protection

• Carcinogenic to Humans

Agency

 Eight epidemiologic studies (ten datasets) in diverse populations used in DRMA

Table 4-5. Pooled meta-regression estimates of extra lifetime lung cancer incidence risk (per 10,000) at various average daily iAs doses using MLE dosesa, b

Extra lifetime	Average daily inorganic arsenic dose (µg/kg-day) ^b											
risk estimates (per 10,000)ª	0	0.02	0.0365 ^b	0.075	0.13	0.185	0.24	0.57	1.12			
5 th percentile	0	0.16	0.29	0.59	1.03	1.46	1.90	4.52	8.90			
Mean	0	3.65	6.67	13.79	24.10	34.57	45.21	112.72	240.40			
95 th percentile	0	8.80	16.15	33.65	59.49	86.35	114.27	306.17	737.83			

^aExtra lifetime risks are presented as mean risk/10,000 with 5%–95% probabilities based on mean, 5% and 95% estimates of dose-response slopes.

^bDoses used in EPA modeling. U.S. daily background dose is estimated at 0.0365 µg/kg, 0.02 µg/kg from diet, $0.0165 \,\mu\text{g/kg}$ from water and $0 \,\mu\text{g/kg}$ from air (see Section 4.3.4).



Lifetime Cancer Slope Factor (CSF): 4.62 $\times 10^{-2} (\mu g/kg-d)^{-1}$

Charge Questions 5d,6

Cancer Dose-Response: Combined CSF

- A combined cancer slope factor was calculated to estimate risk of developing either bladder cancer or lung cancer separately, or both cancers combined
 - Individual tumor slope factors were assumed to be normally distributed
 - Combined CSF equaled the summed central tendency slopes + 1.645 * composite SD

Health Outcome	Hazard Descriptor	Cancer Slope factor (CSF) 1/(µg/kg-day) ^{a, b}		
Bladder cancer		1.3E-2		
Lung cancer	Accepted hazard	4.6E-2		
Combined cancer risk		5.3E-2 ^c		

Table ES-1. Toxicity values for cancer outcomes associated with inorganic arsenic exposure

Noncancer Dose-Response Analyses

- Diseases of the circulatory system (DCS) modeled using Bayesian DRMA approach:
 - Cardiovascular disease (CVD) incidence: two studies
 - Ischemic heart disease (IHD) incidence: four studies
- Diabetes modeled using Bayesian DRMA approach: four studies
- Pregnancy outcomes (birth weight)
 - Single study: Kile et al. (2016)
- Neurodevelopmental effects not advanced primarily due to nonmonotonicity in the exposure-response relationship in critical study (Wasserman et al., (2014)) (see Appendix C.2)

Non-Cancer Dose-Response: DCS

• Evidence demonstrates that iAs causes diseases of the circulatory system in humans based on *robust* evidence in humans

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Table 4-10. Pooled meta-regression estimates of extra lifetime incidence risk (per 10,000) for DCS outcomes at various average daily iAs doses and estimated U.S. equivalent drinking water exposures above median U.S. doses and exposures using MLE doses ^{a, b}

	Extra			Average	daily inor	ganic arse	nic dose (µg/kg-d)¢		
Health outcome	lifetime risk (per 10,000)	0	0.02	0.0365 ^c	0.075	0.13	0.185	0.24	0.57	1.12
CVD incidence	5%	0.00	0.31	0.57	1.18	2.04	2.90	3.77	8.95	17.58
	mean	0.00	32.16	58.65	120.31	208.00	295.24	382.03	892.96	1706.02
	95%	0.00	107.90	196.41	401.18	689.32	972.20	1249.70	2797.36	4918.66
IHD incidence	5%	0.00	0.55	1.00	2.05	3.55	5.05	6.55	15.55	38.39
	mean	0.00	27.34	49.93	102.74	178.41	254.34	330.53	792.25	1687.70
	95%	0.00	70.71	129.24	266.49	464.05	663.18	863.64	2081.04	4539.95
Fatal CVD	5%	0.00	0.11	0.20	0.42	0.72	1.03	1.34	3.17	6.24
	mean	0.00	7.83	14.31	29.50	51.41	73.54	95.89	234.91	485.77
	95%	0.00	22.81	41.82	86.85	152.87	220.89	290.97	756.19	1712.15
Fatal IHD	5%	0.00	0.08	0.15	0.32	0.55	0.78	1.01	2.40	4.72
	mean	0.00	6.59	12.07	24.99	43.76	62.16	82.50	208.70	455.88
	95%	0.00	17.75	32.68	68.51	122.23	173.16	239.15	676.91	1770.18



Charge Question 5a

Lifetime extra risk plots for CVD incidence 22

Non-Cancer Dose-Response: Diabetes

• Evidence demonstrates that iAs causes diabetes in humans based on *robust* evidence in humans

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Table 4-12. Pooled meta-regression estimates of extra lifetime diabetes incidence risk (per 10,000) at various average daily iAs doses and U.S. equivalent drinking water above median U.S. doses and exposures using MLE dose estimates ^{a, b}

Extra lifetime risk		g/kg-d) ^c								
estimates (per 10,000)	0	0 0.02 0.0365 0.075 0.13 0.185 0.24 0.57								
5 th percentile	0	1.72	3.14	6.45	11.18	15.91	20.64	49.06	96.51	
Mean	0	27.49	50.20	103.30	179.38	255.73	332.34	796.62	1581.96	
95 th percentile	0	70.62	129.08	266.14	463.45	662.31	862.51	2078.32	4050.39	



Lifetime extra risk plots for diabetes

SEPA United States Environmental Protection Agency Refence Dose (RfD) Derivations

- Dose-response was harmonized between cancer and non-cancer endpoints were possible
 - Probabilistic risk-at-a-dose values were provided for bladder cancer, lung cancer, DCS endpoints, and diabetes
- Reference doses (RfDs) were additionally provided in the assessment to met the needs of EPA Program and Regional offices
- Definition for the RfD:
 - An estimate (with uncertainty spanning perhaps an order of magnitude) of a daily oral exposure to the human population (including sensitive subgroups) that is likely to be without an appreciable risk of deleterious effects during a lifetime.

SEPA United States Environmental Protection Agency RfD Derivation: CVD incidence

- Using CVD incidence as an example:
 - Lifetime background rate of disease = 0.70 at an assumed background exposure of 0.0365 μg/kg-day iAs
 - Using logistic slope (β mean) = 0.23, probability of response at zero dose, P(0), was estimated to be 0.698
 - Given an extra risk of 5%, P(d) estimated to be 0.713
 - Odds (i.e., $\frac{p}{1-p}$) at P(0) and P(d) calculated as 2.313 and 2.488 and the ratio as 1.075
 - $BMD = \frac{\ln(1.075)}{0.23} = 0.315 \,\mu\text{g/kg-day iAs}$
 - BMDL = 0.094 μ g/kg-day iAs (calculated as above using 95th lower bound on β mean)

EPA United States Environmental Protection Agency RfD Derivation: Birth Weight

- Kile et al. (2016) provided linear regression coefficient for the association between iAs in drinking water and low birth weight: -19.2 g (95% CI: -24.6,-13.7) per ln(µg/L) iAs in drinking water
- This regression coefficent was re-expressed in terms of per μ g/L: -4.3 g (95% CI: -5.5, -3.1) per μ g/L iAs in drinking water
- Then, given the average birth weight in the United States and the percentage of those births falling below the clinical definition of low birth weight (2500 g), the BMR was defined as 5% extra risk of falling below that cut-off
- Rearranging linear equation and solving for dose results in a BMD of 21.4 $\mu g/L$ and a BMDL of 17.3 $\mu g/L$



• DCS and Diabetes:

candidate values based on BMDL₀₅s derived from the 95th upper bound on DRMA logistic slopes

 Pregnancy outcomes (birth weight): candidate value based on BMDL₀₅ derived from the 95th upper bound on studyreported linear regression coefficient Table 4-17. Points of departure (PODs) considered for use in deriving candidate RfDs for iAs

Health outcome	Study	Basis for point of departure	Point of departure (µg/kg-day)						
CVD incidence	Meta-regression of 2 studies	BMDLos	0.094ª						
IHD incidence	Meta-regression of 4 studies	BMDLos	0.128ª						
Diabetes	Meta-regression of 4 studies	BMDL ₀₅	0.127ª						
Birth weight	Kile et al., 2016	BMDLos ^b	0.23						
ln(^{odds}	$\ln(odds at P(d)/sds at P(d))$								

 $^{BMDL} = \frac{\beta MDL}{95^{ch}upper\ bound\ on\ mean(\beta\ mean)}}$, where P(d) and P(0) are the probabilities associated with 5% and 0% extra risk, respectively, see details and modeling results in Appendix C, Section C.3.

Table 4-19. Candidate RfD values for inorganic arsenic (iAs)

Endpoint	POD (µg/kg-d)	UFA	UFH	UFs	UFL	UF₀	UFc	Candidate value (µg/kg-d)
CVD Incidence	0.094	1	3	1	1	1	3	0.031
IHD Incidence	0.128	1	3	1	1	1	3	0.043
Diabetes	0.127	1	3	1	1	1	3	0.042
Birth weight	0.23	1	3	1	1	1	3	0.077

Overall RfD of 0.031 $\mu g/kg$ -day based on CVD incidence



Thank you!

To the Scientific Advisory Board Inorganic Arsenic Panel and staff

Questions?



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